Contents lists available at ScienceDirect





journal homepage: www.elsevier.com/locate/scitotenv

Hazardous cyanobacteria integrity response to velocity gradient and powdered activated carbon in water treatment plants



Samylla Oliveira ^a, Allan Clemente ^a, Indira Menezes ^a, Amanda Gois ^a, Ismael Carloto ^b, Linda Lawton ^c, Jose Capelo-Neto ^{a,*}

^a Federal University of Ceara, Department of Hydraulic and Environmental Engineering, Block 713, Campus Pici, Fortaleza, Ceará, Brazil

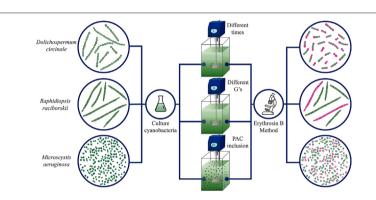
^b Federal Institute of Education, Science, and Technology of Ceará - Campus Acaraú, Av. Des. Armando de Sales Louzada, s/n, Acaraú, Ceará, Brazil

^c Robert Gordon University, Sir Ian Wood Building, Garthdee Road, Aberdeen, United Kingdom

HIGHLIGHTS

GRAPHICAL ABSTRACT

- *D. circinale* total metabolite concentration should be used to estimate PAC dosage.
- The reservoir's dominant species should guide the operations in WTP.
- *D. circinale* and *R. raciborskii* were the most and least sensitive cells in the experiment, respectively.



ARTICLE INFO

Article history: Received 7 July 2020 Received in revised form 21 December 2020 Accepted 7 January 2021 Available online 4 February 2021

Editor: Paola Verlicchi

Keywords: Microcystis aeruginosa, Raphidiopsis raciborskii, and Dolichospermum circinale Drinking water treatment Flash mixing Shear stress Cell integrity

ABSTRACT

Although some studies have investigated the impact caused by chemicals used on water treatment (coagulants and oxidants) on cyanobacteria integrity, the isolated effect of shear stress during coagulation is still not fully understood. This study evaluated the impact of different velocity gradients, mixing times, and the addition of powdered activated carbon (PAC) on the integrity of *Microcystis aeruginosa*, *Raphidiopsis raciborskii*, and *Dolichospermum circinale*, known producers of toxin and taste and odor (T&O) compounds. No association was found between *R. raciborskii* cell lysis and velocity gradient, with or without PAC, demonstrating the high resilience of this taxon to shear stress. In contrast, an association was found for *M. aeruginosa* at the highest velocity gradient evaluated (1000 s^{-1}) and for *D. circinale* above the lowest velocity gradient studied (600 s^{-1}) . After PAC addition, there was a reduction in the chances of finding *M. aeruginosa* intact cells above velocity gradient 80 s⁻¹ at 45 s, while *D. circinale* show cell lysis in all the scenarios expect at 600 s⁻¹ and 10 s of agitation. The additional impact of PAC on cell lysis may lead to more release of metabolites and shows the need to adjust the hydraulic conditions in the rapid mixing stage, especially when more "fragile" cyanobacteria are present. Neither cyanobacterial cell size nor morphology was shown to be relevant to shear stress sensitivity, indicating that cell wall composition might have been an important factor in controlling cell lysis.

© 2021 Published by Elsevier B.V.

1. Introduction

* Corresponding author.

E-mail addresses: allan.clemente@alu.ufc.br (A. Clemente), ismael.lopes@ifce.edu.br (I. Carloto), l.lawton@rgu.ac.uk (L. Lawton), capelo@ufc.br (J. Capelo-Neto).

The rapid anthropic eutrophication aggravated by climate change has made cyanobacteria blooms more frequent and intense, challenging existing water treatment plant (WTP) processes to effectively produce drinking water that is safe and aesthetically acceptable (Bullerjahn et al., 2016; Duran-Encalada et al., 2017; Paerl et al., 2016a; Trenberth et al., 2014). In tropical regions, for example, toxin-producing filamentous cyanobacteria, among them *R. raciborskii*, have been dominating the phytoplankton community of artificial reservoirs, with average cell concentrations of around 10⁶ cells mL⁻¹. *D. circinale* and *M. aeruginosa* have also been of major concern to water companies in these regions (Barros et al., 2017).

Conventional treatments (e.g., coagulation, flocculation, sedimentation, filtration, and disinfection) have multiple barriers that aim to remove colloids, pathogens, and suspended particles, such as cyanobacteria but are generally not designed to remove dissolved compounds (toxins and T&O compounds). In addition to the cell lysis that occurs naturally in the reservoir, several physical and chemical stresses may cause damage to the cyanobacteria membrane throughout the treatment steps, thereby increasing the solubilization of their intracellular content (Szlag et al., 2015; Zamyadi et al., 2015). In this case, WTPs that use water from highly eutrophic reservoirs may produce drinking water with concentrations of toxins and T&O compounds above guideline thresholds, leading to the rejection of the finished water due to the unpleasant organoleptic characteristics or to the potential toxicity (Chernova et al., 2017, 2019; Paerl et al., 2016a). Additionally, the increase of soluble organic matter and secondary metabolites increases WTP chemical requirements and the potential to form disinfection byproducts during the treatment process (Ho et al., 2012). Optimization of the WTP process to remove intact cyanobacterial cells should significantly reduce the costs and risks involved in the treatment process (Sun et al., 2012; Zamyadi et al., 2015).

Different species of cyanobacteria require different strategies during water treatment operations. According to Li et al. (2018), due to greater sensitivity to chemical stresses, *R. raciborskii* and *D. circinale* showed a higher chance of metabolite release during coagulation and sludge storage when compared to *M. aeruginosa*. Pestana et al. (2019) studied three full-scale WTPs and also found that cyanobacteria trichome damage was species-dependent. Cyanobacteria with larger trichome length (*Planktothrix* sp., *Geitlerinema* sp. and *Dolichospermum* sp.) were more prone to breakage than short-trichomed ones (*Pseudanabaena* sp. and *Planktolyngbya* sp.). Clemente et al. (2020) evaluated the cellular and trichome integrity of two filamentous cyanobacteria during the rapidmixing step and found that *D. circinale* was significantly affected while *R. raciborskii* showed no signs of damage.

Although few researchers have studied the effect of coagulants and oxidants on the integrity of cyanobacteria (Bakheet et al., 2018; Naceradska et al., 2017; Xie et al., 2016; Zhou et al., 2014), the isolated impact of the shear stress due to hydraulic conditions applied during the coagulation step on the cell integrity of *R. raciborskii*, *D. circinale*, and *M. aeruginosa* has not been fully understood. Rodríguez et al. (2009) reported that excessive shear stress can cause reduced cell growth and productivity, severe cell damage, and even cell lysis, which could be particularly challenging for sensitive species.

Sensitivity to shear differs between species and is dependent on multiple factors including cell size, presence of flagella, morphology, and composition of the cell wall (Rodriguez et al., 2009). However, contradictory results in the literature highlight the complexity of microalgal behavior and that systematic studies under controlled conditions are still needed to gain a thorough understanding of the effects of the competing parameters (Norsker et al., 2011).

The proposed hypothesis is that shear stress, imposed by the velocity gradient necessary for coagulation, can impose integrity loss on cyanobacteria cells and that the higher the velocity gradient and the longer the organisms are subjected to it, the more likely organisms are to be damaged. Moreover, powdered activated carbon (PAC), added during the coagulation step to remove dissolved secondary metabolites, may increase the chances of cyanobacteria cell damage due to the mechanical stress they add to the suspension during flash mixing. The objective of this study was to evaluate the impact of velocity gradient, agitation time, and PAC addition on *R. raciborskii*, *D. circinale*, and *M. aeruginosa* cell integrity in order to identify which of these variables could be managed so that organism damage and intracellular-content release could be minimized.

2. Materials and methods

2.1. Cyanobacterial cultures

Three cyanobacteria, *R. raciborskii*, *D. circinale*, and *M. aeruginosa* were selected based on their presence in surface reservoirs in several countries and on their potential to produce undesired secondary metabolites (Barros et al., 2017; Haakonsson et al., 2017; Sinha et al., 2012).

The species were isolated from surface reservoirs in Northeast Brazil and cultured in ASM-1 medium (Gorham et al., 1964) with pH 8, in nonaxenic conditions, but with imperceptible bacterial contamination under conventional microscopic analysis. The species were kept under a white light intensity of $6.75 \,\mu$ ·mol·m⁻²·s⁻¹ (Digital Light Meter YF-1065, China), temperature of 24 ± 2 °C, and photoperiod of 12:12 h (light/dark) for 21 ± 2 days. *M. aeruginosa* grew in unicellular spherical form and cells with an average diameter of 2.5 µm. *R. raciborskii* was composed of chained filaments, known as trichomes, with an average length of 75 µm, while *D. circinale* also exhibited straight filamentous morphology but with an average length of 15 µm, both without mucilaginous sheaths.

2.2. Cell density and integrity analysis

Cell density was measured according to APHA-AWWA-WEF (2005) using an inverted optical microscope (Zeiss Vert.A1, Germany) and a Sedgewick-Rafter chamber. Counting was performed across bands or fields with a magnification of $20 \times$ or $40 \times$, allowing a confidence interval of $95\% \pm 20\%$. To quantify the number of cells per trichome and determine the length of filaments, 30 filaments were randomly chosen and evaluated immediately after each experiment using an optical microscope (Olympus Optical, Model: Cx-31, Brazil). Cell counting was performed immediately before the experiments, therefore, no stain or fixative was added to the samples.

Cell integrity was evaluated by a staining method using 5% erythrosine-B (Dynamics, Brazil) by Clemente et al. (2020) with a magnification of $40\times$. Erythrosine-B ($C_{20}H_6I_4Na_2O_5$) is a non-toxic, biological dye that can be used to identify damaged living cells (Clemente et al., 2020; Calomeni and Rodgers, 2015; DiBartolomeis and Mone, 2004; Markelova et al., 2000). In a cell with an intact membrane, erythrosine-B is not absorbed and maintains its original appearance. On the other hand, in cells with compromised integrity, erythrosine-B penetrates, and accumulates in the cytoplasm, producing a pink color easily distinguished under optical microscopy (Markelova et al., 2000). In our experiment, 100 cells of *M. aeruginosa* and *D. circinale* and 100 trichomes of *R. raciborskii* were randomly selected and counted in triplicate to improve the statistical power of each sample. In total, for each species, 10,800 organisms were evaluated totaling 32,400 cell integrity results.

2.3. Experimental design

Each strain of *M. aeruginosa*, *D. circinale*, and *R. raciborskii* were separately cultured in ASM-1 medium and diluted with dechlorinated treated water (Supplementary materials contains the basic water quality data) until the final concentration of 10^6 cells mL⁻¹ (Fan et al., 2014; Qi et al., 2016; Zhou et al., 2014). Continuous stirred reactors were used to evaluate the impact of different velocity gradients and mixing times on the cellular structure of the selected species. The velocity gradients applied were 600 s⁻¹ (270 rpm), 800 s⁻¹ (345 rpm), and 1000 s⁻¹ (420 rpm) at mixing times of 10, 15, 30, 45, and 60 s. These parameters

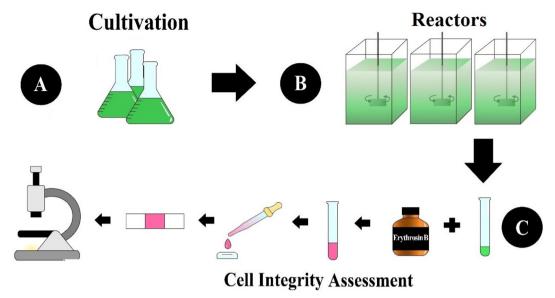


Fig. 1. Simplified experimental protocol. (A) represents the species cultivation; (B) shows the experiments using different velocity gradients, mixing times, and PAC addition; (C) represents the identification of cell integrity.

were based on the coagulation conditions (Velocity Gradient 600–1000 s⁻¹; Mixing times 10–60 s) suggested by AWWA and ASCE (2012). All experiments were performed in triplicate.

In the second phase of the experiment, PAC (Coconut-based; bulk density = 425 kg m⁻³; mean particle diameter = 30 μ m) was added to the reactors under the same conditions applied in the previous phase, to observe the impact of carbon particles on the cyanobacterial cell structure. A PAC dose of 30 mg L⁻¹ was applied since it is commonly used in both the experimental and full-scale (Kim et al., 2014). The simplified experimental protocol is presented in Fig. 1.

2.4. Data analyses

Statistical analyses to evaluate the hypothesis of the association between the variables (cyanobacteria species, velocity gradient interval, mixing time, and with or without PAC addition) and cell integrity was performed using *RStudio* and modified data analysis methodology (Clemente et al., 2020). Details on these data analyses can be found in the Supplementary materials.

3. Results and discussion

3.1. Identifying association between cell integrity and experimental variables

Tables 1, 2, and 3 present the mean and standard deviation of intact cyanobacteria cells after exposure to velocity gradients G1 (600 s^{-1}), G2 (800 s^{-1}) and G3 (1000 s^{-1}), with and without added PAC and at mixing times of 10, 15, 30, 45 and 60 s. In these tables, the results of Fisher's exact test were incorporated, to compare the impacts of different mixing times to the initial condition (t = 0 s). The association between cell integrity and any variable was considered significant when p < 0.05.

R. raciborskii cell integrity was not compromised in any experimental condition (Table 1), even with the most intense velocity gradient, highest mixing time, and after PAC addition. Therefore, there is no evidence that mechanical stress imposed by the hydraulic condition usually applied in WTPs is capable of compromising *R. raciborskii* cells. The cellular integrity of *M. aeruginosa* (Table 2) was only affected by intense velocity gradient (G3 [1000] s⁻¹), without PAC, and at relatively long times (45 and 60 s). Reductions of up to 4% in the number of intact cells in the other scenarios were not considered significant. Table 3

shows sufficient statistical evidence that, without PAC, velocity gradients above G2 ([800] s⁻¹) may cause a reduction in the number of *D. circinale* intact cells. With the introduction of PAC, starting from G1 ([600] s⁻¹) and mixing time of 15 s, there was a sharp drop in the number of *D. circinale* intact cells. The results show that *D. circinale* is more sensitive to mechanical stress caused by the hydraulic conditions than *R. raciborskii* and *M. aeruginosa* and that the presence of PAC intensifies this effect.

Shear stress has been shown to cause filament breakage in different filamentous cyanobacteria. *Spirulina platensis* and *Aphanizomenon flosaquae*, for example, displayed shear sensitivity compared to green algae and other cyanobacteria (Bronnenmeier and Märkl, 1982; Leupold et al., 2013). Our results contradict the Wang and Lan (2018) study which gathered enough evidence to suggest that filamentous phytoplankton displayed greater shear sensitivity than unicellular ones.

Although it has been hypothesized that larger microalgae are possibly more sensitive to shear stress than smaller ones (Camacho et al.,

Table 1

Impact of mixing conditions and PAC on R raciborskii cell integrity (mean # intact cells/ 100, SD, Fisher's statistic).

Raphidiopsis raciborskii										
PAC	Velocity		Time of mix							
	gradient (s ⁻¹)		0 s	10 s	15 s	30 s	45 s	60 s		
Without	600	Mean (SD) P	100 (0.00)	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1		
	800	Mean (SD) P	100 (0.00) -	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1		
	1000	Mean (SD) P	100 (0.00) -	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1		
With	600	Mean (SD) P	100 (0.00) -	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1		
	800	Mean (SD) P	100 (0.00) _	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1		
	1000	Mean (SD) P	100 (0.00) -	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1	100 (0.00) 1		

Table 2

Impact of mixing conditions and PAC on *M. aeruginosa* cell integrity (mean # intact cells/ 100, SD, Fisher's statistic).

PAC	Velocity		Time of mix							
	gradient (s ⁻¹)		0 s	10 s	15 s	30 s	45 s	60 s		
Without	600	Mean (SD) P	100 (0.58) -	99 (0.58) 1	100 (0.00) 1	97 (0.58) 0.246	97 (0.58) 0.246	97 (0.58) 0.246		
	800	Mean (SD) P	99 (1.15) -	98 (1.15) 1	98 (1.53) 1	97 (2.65) 1	97 (0.58) 1	96 (0.58) 0.369		
	1000	Mean (SD) P	99 (0.58) -	96 (1.15) 0.369	94 (1.53) 0.120	94 (0.00) 0.120	91 (2.52) 0.018	92 (2.08) 0.035		
With	600	Mean (SD) P	100 (0.58)	100 (0.00) 1	100 (0.58) 1	99 (0.00) 1	97 (1.53) 0.246	96 (0.58) 0.121		
	800	Mean (SD) P	100 (0.58)	99 (0.58) 1	96 (0.58) 0.121	96 (1.53) 0.121	97 (1.73) 0.246	96 (1.53) 0.121		
	1000	Mean (SD) P	100 (0.58) -	99 (0.00) 1	99 (0.00) 1	98 (0.58) 0.497	97 (1.00) 0.246	96 (1.00 0.121		

2001), this was also not true in our study since the largest cyanobacteria in our study, *R. raciborskii*, was the least affected by shear stress.

Cell walls provide structural support, protective covering, and play a role in shear sensitivity (Barbosa et al., 2003). Most green algae contain a cellulose-rich cell wall, which provides a rigid and resilient covering, but cyanobacterial cell walls normally consist of three to four layers mainly composed of peptidoglycan (Gumbo et al., 2008). Interestingly, green algae *Chlamydomonas reinhardtii* is cellulose-deficient and quite sensitive to shear stress (Jeanneret et al., 2016).

M. aeruginosa cell wall is divided into several layers: the cytoplasmic membrane, the peptidoglycan layer, and the multilayered structure of the cell wall composed of polysaccharides, which lend rigidity to the cell wall (Woitzik et al., 1988). The literature contains a very limited amount of information about the cell wall composition of *D. circinale* or *R. raciborskii.*

As a proxy to *D. circinale* cell wall composition, we examined the composition of *Anabaena cylindrica* cells as presented by Dunn and Wolk (1970). The vegetative cell wall contained amino compounds (65%),

Table 3

Impact of mixing conditions and PAC on D. circinale cell integrity (mean # intact cells/100, SD, Fisher's statistic).

Dolichospermum circinale										
РАС	Velocity		Time of mix							
	gradient (s ⁻¹)		0 s	10 s	15 s	30 s	45 s	60 s		
Without	600	Mean	80	77	84	80	85	74		
		(SD)	(8.62)	(7.64)	(6.66)	(8.14)	(14.30)	(1.15)		
		р	-	0.731	0.581	1	0.457	0.401		
	800	Mean	62	39	49	60	40	41		
		(SD)	(2.52)	(14.40)	(16.92)	(17.24)	(8.89)	(3.51)		
		р	-	0.002	0.088	0.885	0.003	0.005		
	1000	Mean	96	93	85	82	84	68		
		(SD)	(3.61)	(4.04)	(3.61)	(1.00)	(4.36)	(8.89)		
		р	-	0.537	0.014	0.002	0.008	<0.001		
With	600	Mean	91	83	75	71	59	53		
		(SD)	(2.08)	(1.73)	(3.79)	(2.00)	(3.21)	(2.65)		
		р	-	0.140	0.004	<0.001	<0.001	<0.001		
	800	Mean	92	81	77	78	72	34		
		(SD)	(5.57)	(9.64)	(2.08)	(2.08)	(2.89)	(6.36)		
		р	_	0.037	0.006	0.009	< 0.001	< 0.001		
	1000	Mean	93	77	76	69	56	45		
		(SD)	(5.57)	(9.64)	(2.08)	(2.08)	(2.89)	(6.36)		
		р	-	0.002	0.001	<0.001	<0.001	< 0.001		

lipid (3%), and polysaccharides (18%). The polysaccharides consisted mainly of mannose, with smaller amounts of glucose, galactose, fucose, and xylose. This relatively low amount of polysaccharides, when compared to the other cyanobacteria analyzed in the present study, may explain why *D. circinale* is the most shear stress-sensitive among the studied taxa (Woitzik et al., 1988). Andrade et al. (2010) observed that brown algae produced up to 49% of polysaccharides in the cell wall as a protection mechanism against chemical stress.

Zarantonello et al. (2018) studied the structure of the *R. raciborskii* cellular envelope at higher magnification and revealed that this structure was composed of two bilayered membranes: the inner or plasma membrane and the outer membrane that encloses the periplasmic space with a thin peptidoglycan layer. Hamilton et al. (2005) also observed that *R. raciborskii* cell walls were notably thick and conspicuous which might explain the highest resistance to the shear stress in our experiments.

In recent years, several researchers have shown that *D. circinale*, R. raciborskii, and M. aeruginosa respond differently to physical and chemical stimuli in WTPs. Chow et al. (1999) demonstrated that, when *M. aeruginosa* was subjected to velocity gradients of up to 500 s^{-1} for 4 h, neither significant changes in cellular integrity nor microcystin-LR release was observed, corroborating with the data on this study, which shows that *M. aeruginosa* cell damage may occur only at high-velocity gradients ($\geq 800 \text{ s}^{-1}$ corresponding to 345 rpm). Sun et al. (2012) observed no M. aeruginosa cell lysis or toxin release at different velocity gradients (40, 150, 250, and 500 rpm), with a mixing time of 20 min, and without the addition of chemicals. It is worth mentioning that in the first three velocity gradients (40, 150, and 250 rpm) used by Sun et al. (2012), the results were similar to those found in the present study. On the other hand, divergence was observed for the highest velocity gradient (500 rpm) since there was a loss of cellular integrity at 1000 s⁻¹ or 420 rpm in the current experiment.

Clemente et al. (2020) evaluated the effects of velocity gradient and mixing time of rapid and slow mixing on *D. circinale* and *R. raciborskii* cells. The rapid-mixing condition (750 s⁻¹ or 325 rpm for 60 s) compromised *D. circinale* cells reducing the number of intact organisms, while *R. raciborskii* was not affected by the same hydraulic condition. Therefore, the authors also observed that *R. raciborskii* and *D. circinale* cells were differently affected by the mechanical stress of hydraulic conditions.

Li et al. (2018) compared R. raciborskii and M. aeruginosa membrane damage during coagulation using polyaluminum ferric chloride (PAFC). The results showed that intact *R. raciborskii* cells were removed under optimal coagulation conditions but, when high dosages (>10 mg L^{-1}) were applied, toxin release increased. According to Li et al. (2018), M. aeruginosa cells only released toxin with PAFC dosages higher than 30 mg L^{-1} , due to morphological and physiological differences. The results presented by Li et al. (2018) corroborate those found in the present study, also demonstrating a distinct resilience of R. raciborskii and M. aeruginosa to shear stress during the treatment process. Mucci et al. (2017) reported different cyanobacterial sensitivities due to the use of chitosan. The less sensitive strains were a member of the Chroococcales order (Microcystis sp.) and one of the Nostocales (Aphanizomenon sp.), while the most sensitive were the Nostocales (Cylindrospermopsis sp.) and the Oscillatorias (Planktothrix sp.). Although the present study used no chemical additives, a similar behavior to the aforementioned studies was observed when evaluating D. circinale, R. raciborskii, and M. aeruginosa cell sensitivities. WTP managers and operators could consider this pattern to guide their operational procedures based on the phytoplankton composition of the raw water.

3.2. Elaboration of the logistic regression model for odds ratio determination

The extent of the damage to cellular integrity was evaluated using logistic regression and was found to be closely associated with gradient

exposure, mixing time, and presence or absence of PAC for *M. aeruginosa* and *D. circinale*. This method was used because it helps evaluate the three variables simultaneously (velocity gradient, mixing time, and PAC addition), including considering their synergisms (Agresti, 2012). A logistic regression model, similar to the one proposed herein, was used by Sheng et al. (2012) to identify the response relationship between cyanobacterial blooms and environmental factors such as water quality and climatic conditions at Dianchi Lake, China. In addition to Sheng et al. (2012), several authors (Aung et al., 2019; Egorov et al., 2018; García-Pérez et al., 2016; Tornevi et al., 2016) have adopted a similar statistical approach, which allows us to infer that this method is appropriate for this study.

Table 4 shows the effects of the variables X_1 (cyanobacterial species), X_2 (PAC addition), X_3 (velocity gradient), X_4 (mixing time) and their double ($X_1.X_3, X_3.X_2, X_1.X_2, X_4.X_3$, and $X_4.X_2$) and triple interactions ($X_1.X_3, X_2$ and $X_4.X_3.X_2$) on cellular integrity. In each different model, a new variable or group of variables was added to the null model M1 and verified if there was a sharp drop in the deviance (p < 0.05) and Akaike Information Criterion (AIC) values. In model M2, the null model M1 inserted with the effect of cyanobacteria species (X_1), presented the most significant reduction of deviance, showing that species have the greatest relevance for predicting the degree of association between the variables and cellular integrity. These results corroborate with Clemente et al. (2020) study, which associated cyanobacterial membrane damage on different cyanobacterial species using the same treatment conditions used in our work.

The addition of the variables x_4 (PAC addition) and x_3 (velocity gradient) into models M3 and M4, respectively, led to significant reductions in deviance, inferring that mixing time and velocity gradient are closely associated with the decrease in cellular integrity. This reinforces the findings by Clemente et al. (2020) who showed that hydraulic conditions (velocity gradient of 750 s⁻¹ or 350 rpm and mixing time of 60 s) compromised *D. circinale* cellular integrity. On the other hand, M5 (M4 + PAC addition - X₂) revealed no significant reduction in deviance and, therefore, the addition of PAC, when analyzed separately, had no impact on cellular integrity.

When considering the synergy effects of velocity gradient, mixing time, and the presence of PAC (M6 through M12), the addition of multiple interactions into the model reduced the deviance more significantly than that of model M5, implying that the impact of PAC is relevant in reducing the cellular integrity only when combined with the other two variables. Model M12, shown in Eq. (2), was selected to estimate the degree of association between the variables and cellular integrity since it had the lowest residual deviance and AIC (Table 4). M12 also simultaneously fulfilled the necessary conditions for Pearson and deviance residuals, which were within their respective simulated envelope graphs (Fig. 2). In addition, it presented QL = 29.29 (p = 0.502) and QP = 26.21 (p = 0.665).

Table 4

Analysis of deviance (ANODEV) and AIC obtained in the cellular integrity study of *M. aeruginosa* and D. circinale. (Mi): fitted model; (X₁) cyanobacteria species, (X₂) PAC addition; (X₃) velocity gradient; and (X₄) mixing time.

Model	Factor	Deviance	Residual deviance	p (chi)	AIC
M1	NULL		1673.06	-	1921.87
M2	$M1 + X_1$	1086.18	586.87	< 0.001	837.69
M3	$M2 + X_4$	208.08	378.79	< 0.001	639.61
M4	$M3 + X_3$	88.37	290.43	< 0.001	555.24
M5	$M4 + X_2$	1.24	289.18	0.265	_
M6	$M5 + X_1.X_3$	28.61	260.57	< 0.001	530.64
M7	$M6 + X_3 X_2$	100.82	159.76	< 0.001	-
M8	$M7 + X_1 X_2$	14.46	145.29	< 0.001	-
M9	$M8 + X_{4.}X_{3}$	22.35	122.95	0.013	528.82
M10	$M9 + X_4 X_2$	42.35	80.60	< 0.001	-
M11	$M10 + X_1.X_3.X_2$	26.82	53.78	< 0.001	-
M12	$M11 + X_4 X_3 X_2$	24.50	29.29	0.006	360.1

$$Model M12\tilde{X}_1 + X_3 + X_4 + X_1 * X_2 * X_3 + X_2 * X_3 * X_4$$
(2)

where $X_1 * X_2 * X_3$ and $X_2 * X_3 * X_4$ represent all possible interactions between variables.

3.3. Measuring the association between cell integrity and experimental variables

The previous results (Table 4) identified an association between higher mixing times and cell lysis. However, due to the samples` variability, the effect of PAC addition on cyanobacterial lysis could not be identified since no significant difference (P > 0,05) in cell damage was observed when comparing organisms subjected to those not subjected to PAC, for all times evaluated. Therefore, to better understand the effects of PAC addition, two scenarios (with and without PAC addition) were compared using the odds ratio (OR). The odds ratio represents the ratio between the odds of an event to occur in one group (i.e. samples with addition of PAC) to the odds of that event to occur in another group (i.e. samples with no addition of PAC). Since estimated logistic regression considers the relation of variables, the estimated odds ratio (\widehat{OR}) represents the best way of measuring the impact of PAC on the cell integrity.

Fig. 3 shows the variation of the odds ratio at times 10, 15, 30, 45, and 60 s in relation to zero time in each experiment of *D. circinale* and *M. aeruginosa*, at G1 [600] s⁻¹, G2 [800] s⁻¹, and G3 [1000] s⁻¹, with and without PAC addition. The graphs in Fig. 3 present two regions separated by a dotted line, which indicates an odds ratio value of one. If the odds ratio and its respective confidence interval (Error bar) fall below the dotted line, the probability of finding intact cells in the presence of PAC is lower than in the absence of PAC. On the other hand, if both the odds ratio and its confidence interval are above the dotted line, the effect of introducing PAC increases the odds of finding intact cells. When the odds ratio is over the dotted line or its confidence interval is crossing it, the odds of finding intact cells with or without PAC are equal (See a more detailed explanation in Supplementary materials).

With the presence of PAC at $G = [600] \text{ s}^{-1}$ (G1), mixing time did not change significantly the odds of finding *M. aeruginosa* intact cells when compared to identical hydraulic conditions in the absence of PAC. However, at $G = [800] \text{ s}^{-1}$ (G2) and mixing time of 60 s, the odds of finding intact cells were lower in the presence of PAC than in the absence of PAC. At $G = [1000] \text{ s}^{-1}$ (G3), the odds of finding *M. aeruginosa* intact cells were reduced at all experimental times. This suggests that cell lysis events in the presence of PAC can be avoided at low-velocity gradients and, if higher gradients are required, mixing time should be reduced.

Analyzing *D. circinale* suspension exposed to $G = [600] s^{-1} (G1)$, the odds of finding *D. circinale* intact cells did not increase with the presence of PAC at 10 s. However, after this mixing time (10 s), the odds of cellular integrity loss were estimated to increase by 50% when compared to the experiment in the absence of PAC. Applying $G = [1000] s^{-1} (G3)$ on *D. circinale*, the odds of finding intact cells were reduced even at the shortest mixing time (10 s). However, no conclusive results could be observed at $G = [800] s^{-1} (G2)$ probably due to the lower number of integral cells (62 ± 2.52) at the beginning of the experiment with PAC (Table 3). Since during the experiments we observed that *D. circinale* cells are extremely sensitive, we believe that this experiment already started with a large number of damaged cells.

In one of the few studies reported in the literature using *Dolichospermum* sp. under similar conditions to the ones used in our study, Moisander et al. (2002) demonstrated that this genus is more susceptible to shear stress than other genera of filamentous cyanobacteria. In that context, although several studies have attributed the impact of WTP on cell integrity damage to the use of chemical compounds, we believe that their experimental designs may have led to an overestimated evaluation of cell lysis blamed on these chemical agents.

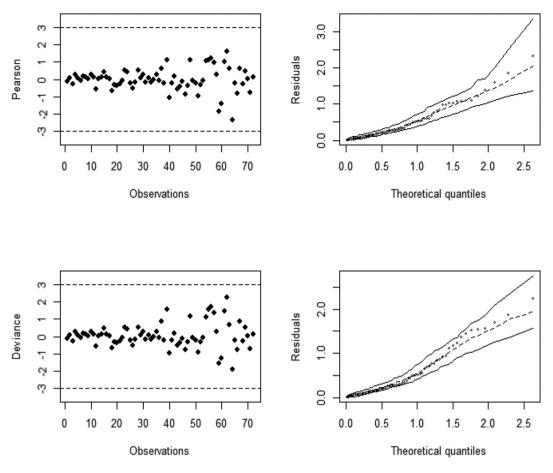


Fig. 2. Pearson and deviance residuals, and their respective simulated envelope graphs.

In another study that also evaluated the effects of hydraulic conditions on cyanobacterial cells, Clemente et al. (2020) observed that in a high-velocity gradient (750 s⁻¹ or 325 rpm) and low mixing time (60 s) the odds of finding intact *D. circinale* were smaller than the odds of finding intact *R. raciborskii*. The odds of finding intact cells of *R. raciborskii* were 124 times greater than *D. circinale*.

Intense *D. circinale* lysis is of major concern to water companies since many authors have reported that this taxon is potentially producers of geosmin, 2-metil-isoborneol (MIB), and cyanotoxins such as saxitoxin, microcystin, cylinderspermopsin, anatoxin-a, and anatoxin-a(S) (Carmichael and Boyer, 2016; Chernova et al., 2019; Li et al., 2016; Paerl et al., 2016b; Pereyra et al., 2017). An important and concerning assessment of our finding is that the amount of adsorbent applied to the treatment could be less than needed since the addition of PAC is commonly estimated based on the concentration of dissolved metabolites (e.g., MIB, Geosmina, or toxin) before raw water enters the WTPs (Cook et al., 2001). In that sense, we believe that PAC should be dosed, in the case of *D. circinale*, considering the total metabolite concentration instead of the dissolved one.

4. Conclusions and recommendations

The mechanical stress imposed by the velocity gradients usually applied in WTPs was unable to compromise *R. raciborskii* cells, while the cellular integrity of *M. aeruginosa* was significantly affected only by intense velocity gradients and at relatively long exposure times. With PAC addition, a low-velocity gradient and low mixing time were enough to significantly reduce the number of *D. circinale* intact cells. The statistical analyses showed that *D. circinale* is more sensitive to

mechanical stress than *R. raciborskii* and *M. aeruginosa* and that the presence of PAC intensifies this effect, indicating that WTP operation strategies should be guided by the relevant species present in the raw water.

One of the most evident limitations of our study is that it was performed using cultured organisms in a controlled environment. Morphological differences between our strains and the ones found in the natural environment may induce a divergent response concerning variation in the working variables.

We believe that the main impact of this study is the practical operational tips or suggestions to water treatment plants operators which includes:

- to decrease the velocity gradient, time, or both during a cyanobacterial bloom, reducing the chances of further solubilizing secondary metabolites, or;
- to increase the dosage of PAC to cope with the higher concentration of toxins and T&O compounds released during the coagulation process.

CRediT authorship contribution statement

Samylla Oliveira: Methodology, Investigation, Data curation, Writing – original draft, Supervision. **Allan Clemente:** Validation, Formal analysis, Data curation. **Indira Menezes:** Methodology, Investigation, Resources. **Amanda Gois:** Methodology, Investigation, Resources. **Ismael Carloto:** Investigation, Data curation. **Linda Lawton:** Writing – review & editing. **Jose Capelo-Neto:** Conceptualization, Methodology, Writing – review & editing, Supervision, Project administration, Funding acquisition.

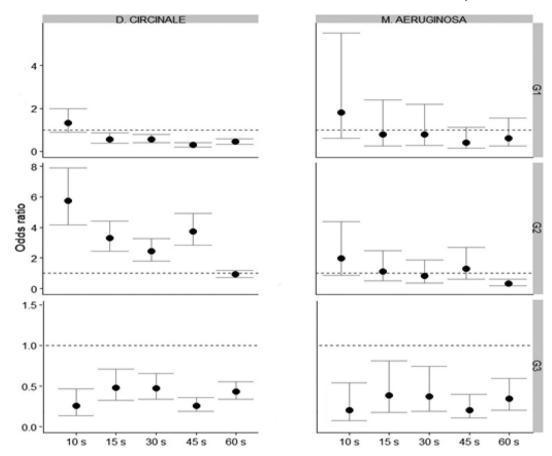


Fig. 3. Impact of velocity gradients (G1, G2, G3) and mixing times on the association between PAC application and cellular integrity (Odds ratio) of D. circinale and M. aeruginosa over 60 s.

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Acknowledgments

The authors would like to thank the Engineering and Physical Sciences Research Council (EPSRC) [EP/P029280/1], the Coordination for the Improvement of Higher Education Personnel - CAPES [PROEX 20/2016 and PrInt 88887.311806/2018-00], the Brazilian National Research Council – CNPq [403116/2016-3 and 304164/2017-8], the Ceará Research Support Foundation - FUNCAP [PNE-0112-00042.01.00/16] for funding this research. Further, the first author also acknowledges the scholarship from the Brazilian National Research Council – CNPq.

Appendix A. Supplementary data

Supplementary data to this article can be found online at https://doi. org/10.1016/j.scitotenv.2021.145110.

References

- Agresti, A., 2012. Categorical Data Analysis. 2nd ed. John Wley & Sons, New Jersey.
- Andrade, L.R., Leal, R.N., Noseda, M., Duarte, M.E., Pereira, M.S., Mourão, P.A., Farina, M., Amado Filho, G.M., 2010. Brown algae overproduce cell wall polysaccharides as a protection mechanism against the heavy metal toxicity. Mar. Pollut. Bull. 60 (9), 1482–1488. https://doi.org/10.1016/j.marpolbul.2010.05.004.
- APHA-AWWA-WEF, 2005. Standard Methods for the Examination of Water and Wastewater. 21th ed. New York.
- Aung, M.T., Ferguson, K.K., Cantonwine, D.E., McElrath, T.F., Meeker, J.D., 2019. Preterm birth in relation to the bisphenol A replacement, bisphenol S, and other phenols

and parabens. Environ. Res. 169, 131–138. https://doi.org/10.1016/j. envres.2018.10.037.

- AWWA, (American Water Works Association), ASCE, (Engineering American Society of Civil), 2012. In: Randtke, S.J. (Ed.), Mixing, Coagulation, and Flocculation.
- Bakheet, B., Islam, M.A., Beardall, J., Zhang, X., McCarthy, D., 2018. Electrochemical inactivation of Cylindrospermopsis raciborskii and removal of the cyanotoxin cylindrospermopsin. J. Hazard. Mater. 344, 241–248. https://doi.org/10.1016/j. jhazmat.2017.10.024.
- Barbosa, M.J., Albrecht, M., Wijffels, R.H., 2003. Hydrodynamic stress and lethal events in sparged microalgae cultures. Biotechnol. Bioeng. 83, 112–120. https://doi.org/ 10.1002/bit.10657.
- Barros, M.U.G., Lopes, I.K.C., Carvalho, S.M. de C., Capelo Neto, J., 2017. Impact of filamentous cyanobacteria on the water quality of two tropical reservoirs. Rbrh. https://doi. org/10.1590/2318-0331.011716072.

Bronnenmeier, R., Märkl, H., 1982. Hydrodynamic stress capacity of microorganisms. Biotechnol. Bioeng. 24, 553–578. https://doi.org/10.1002/bit.260240304.

- Bullerjahn, G.S., McKay, R.M., Davis, T.W., Baker, D.B., Boyer, G.L., D'Anglada, L.V., Doucette, G.J., Ho, J.C., Irwin, E.G., Kling, C.L., Kudela, R.M., Kurmayer, R., Michalak, A.M., Ortiz, J.D., Otten, T.G., Paerl, H.W., Qin, B., Sohngen, B.L., Stumpf, R.P., Visser, P.M., Wilhelm, S.W., 2016. Global solutions to regional problems: collecting global expertise to address the problem of harmful cyanobacterial blooms. A Lake Erie case study. Harmful Algae 54, 223–238. https://doi.org/10.1016/j.hal.2016.01.003.
- Calomeni, A.J., Rodgers, J.H., 2015. Evaluation of the utility of six measures for algal (Microcystis aeruginosa, Planktothrix agardhii, and Pseudokirchneriella subcapitata) viability. Ecotoxicol. Environ. Saf. 111, 192–198. https://doi.org/10.1016/j. ecoenv.2014.09.033.
- Camacho, F.G., Grima, E.M., Mirón, A.S., Pascual, V.G., Chisti, Y., 2001. Carboxymethyl cellulose protects algal cells against hydrodynamic stress. Enzym. Microb. Technol. 29, 602–610. https://doi.org/10.1016/S0141-0229(01)00442-2.
- Carmichael, W.W., Boyer, G.L. 2016. Health impacts from cyanobacteria harmful algae blooms: implications for the North American Great Lakes. Harmful Algae 54, 194–212. https://doi.org/10.1016/j.hal.2016.02.002.
- Chernova, E., Sidelev, S., Russkikh, I., Voyakina, E., Babanazarova, O., Romanov, R., Kotovshchikov, A., Mazur-Marzec, H., 2017. Dolichospermum and Aphanizomenon as neurotoxins producers in some Russian freshwaters. Toxicon 130, 47–55. https:// doi.org/10.1016/j.toxicon.2017.02.016.
- Chernova, E., Sidelev, S., Russkikh, I., Voyakina, E., Zhakovskaya, Z., 2019. First observation of microcystin- and anatoxin-a-producing cyanobacteria in the easternmost part of the Gulf of Finland (the Baltic Sea). Toxicon 157, 18–24. https://doi.org/10.1016/j. toxicon.2018.11.005.
- Chow, C.W.K., Drikas, M., House, J., Burch, M.D., Velzeboer, R.M.A., 1999. The impact of conventional water treatment processes on cells of the cyanobacterium Microcystis

Science of the Total Environment 773 (2021) 145110

aeruginosa. Water Res. 33, 3253-3262. https://doi.org/10.1016/S0043-1354(99) 00051-2.

- Clemente, A., Wilson, A., Oliveira, S., Menezes, I., Gois, A., Capelo-Neto, J., 2020. The role of hydraulic conditions of coagulation and flocculation on the damage of cyanobacteria. Sci. Total Environ. 740, 139737.
- Cook, D., Newcombe, G., Sztajnbok, P., 2001. The application of powdered activated carbon for MIB and geosmin removal: predicting PAC doses in four raw waters. Water Res. https://doi.org/10.1016/S0043-1354(00)00363-8.
- DiBartolomeis, S.M., Mone, J.P., 2004. Apoptosis: a four-week laboratory investigation for advanced molecular and cellular biology students. Cell Biol. Educ. 2, 275–295. https:// doi.org/10.1187/cbe.03-06-0027.
- Dunn, J.H., Wolk, C.P., 1970. Composition of the cellular envelopes of Anabaena cylindrica. J. Bacteriol. 103, 153–158.
- Duran-Encalada, J.A., Paucar-Caceres, A., Bandala, E.R., Wright, G.H., 2017. The impact of global climate change on water quantity and quality: a system dynamics approach to the US–Mexican transborder region. Eur. J. Oper. Res. 256, 567–581. https://doi. org/10.1016/j.ejor.2016.06.016.
- Egorov, A.I., Griffin, S.M., Ward, H.D., Reilly, K., Fout, G.S., Wade, T.J., 2018. Application of a salivary immunoassay in a prospective community study of waterborne infections. Water Res. 142, 289–300. https://doi.org/10.1016/j.watres.2018.05.030.
- Fan, J., Hobson, P., Ho, L., Daly, R., Brookes, J., 2014. The effects of various control and water treatment processes on the membrane integrity and toxin fate of cyanobacteria. J. Hazard. Mater. 264, 313–322. https://doi.org/10.1016/j.jhazmat.2013.10.059.
- García-Pérez, J., Morales-Piga, A., Gómez-Barroso, D., Tamayo-Uria, I., Pardo Romaguera, E., Fernández-Navarro, P., López-Abente, G., Ramis, R., 2016. Risk of neuroblastoma and residential proximity to industrial and urban sites: a case-control study. Environ. Int. 92–93, 269–275. https://doi.org/10.1016/j.envint.2016.04.023.
- Gorham, P.R., McLachlan, J., Hammer, U.T., Kim, W.K., 1964. Isolation and culture of toxic strains of Anabaena flos-aquae (Lyngb.) de Bréb. SIL Proceedings, 1922–2010. vol. 15, pp. 796–804. https://doi.org/10.1080/03680770.1962.11895606.
- Gumbo, R., Ross, G., Cloete, E., 2008. Biological control of Microcystis dominated harmful algal blooms. Afr. J. Biotechnol. 7, 4765–4773.
- Haakonsson, S., Rodríguez-Gallego, L., Somma, A., Bonilla, S., 2017. Temperature and precipitation shape the distribution of harmful cyanobacteria in subtropical lotic and lentic ecosystems. Sci. Total Environ. 609, 1132–1139. https://doi.org/10.1016/j. scitotenv.2017.07.067.
- Hamilton, P.B., Ley, L.M., Dean, S., Pick, F.R., 2005. The occurrence of the cyanobacterium Cylindrospermopsis raciborskii in Constance Lake: an exotic cyanoprokaryote new to Canada. Phycologia 44 (1), 17–25.
- Ho, L., Dreyfus, J., Boyer, J., Lowe, T., Bustamante, H., Duker, P., Meli, T., Newcombe, G., 2012. Fate of cyanobacteria and their metabolites during water treatment sludge management processes. Sci. Total Environ. 424, 232–238. https://doi.org/10.1016/j. scitotenv.2012.02.025.
- Jeanneret, R., Contino, M., Polin, M., 2016. A brief introduction to the model microswimmer Chlamydomonas reinhardtii. Eur. Phys. J. Spec. Top. 225, 2141–2156. https://doi.org/10.1140/epjst/e2016-60065-3.
- Kim, C., Lee, S. II, Hwang, S., Cho, M., Kim, H.S., Noh, S.H., 2014. Removal of geosmin and 2methylisoboneol (2-MIB) by membrane system combined with powdered activated carbon (PAC) for drinking water treatment. J. Water Process Eng. 4, 91–98. https:// doi.org/10.1016/j.jwpe.2014.09.006.
- Leupold, M., Hindersin, S., Gust, G., Kerner, M., Hanelt, D., 2013. Influence of mixing and shear stress on Chlorella vulgaris, Scenedesmus obliquus, and Chlamydomonas reinhardtii. J. Appl. Phycol. 25, 485–495. https://doi.org/10.1007/s10811-012-9882-5.
- Li, X., Dreher, T.W., Li, R., 2016. An overview of diversity, occurrence, genetics and toxin production of bloom-forming Dolichospermum (Anabaena) species. Harmful Algae 54, 54–68. https://doi.org/10.1016/j.hal.2015.10.015.
- Li, H., Pei, H., Xu, H., Jin, Y., Sun, J., 2018. Behavior of Cylindrospermopsis raciborskii during coagulation and sludge storage – higher potential risk of toxin release than Microcystis aeruginosa? J. Hazard. Mater. 347, 307–316. https://doi.org/10.1016/j. jhazmat.2018.01.009.
- Markelova, A.G., Vladimirova, M.G., Kuptsova, E.S., 2000. A comparison of cytochemical methods for the rapid evaluation of microalgal viability. Russ. J. Plant Physiol. 47, 815–819. https://doi.org/10.1023/A:1026619514661.
- Moisander, P.H., Hench, J.L., Kononen, K., Paerl, H.W., 2002. Small-scale shear effects on heterocystous cyanobacteria. Limnol. Oceanogr. 47, 108–119. https://doi.org/ 10.4319/lo.2002.47.1.0108.
- Mucci, M., Noyma, N.P., de Magalhães, L., Miranda, M., van Oosterhout, F., Guedes, I.A., Huszar, V.L.M., Marinho, M.M., Lürling, M., 2017. Chitosan as coagulant on

cyanobacteria in lake restoration management may cause rapid cell lysis. Water Res. 118, 121–130. https://doi.org/10.1016/j.watres.2017.04.020.

- Naceradska, J., Pivokonsky, M., Pivokonska, L., Baresova, M., Henderson, R.K., Zamyadi, A., Janda, V., 2017. The impact of pre-oxidation with potassium permanganate on cyanobacterial organic matter removal by coagulation. Water Res. https://doi.org/ 10.1016/j.watres.2017.02.029.
- Norsker, N.-H., Barbosa, M.J., Vermuë, M.H., Wijffels, R.H., 2011. Microalgal production a close look at the economics. Biotechnol. Adv. 29, 24–27. https://doi.org/10.1016/j. biotechadv.2010.08.005.
- Paerl, H.W., Gardner, W.S., Havens, K.E., Joyner, A.R., McCarthy, M.J., Newell, S.E., Qin, B., Scott, J.T., 2016a. Mitigating cyanobacterial harmful algal blooms in aquatic ecosystems impacted by climate change and anthropogenic nutrients. Harmful Algae 54, 213–222. https://doi.org/10.1016/j.hal.2015.09.009.
- Paerl, H.W., Qin, B., McCarthy, M.J., Newell, S.E., Havens, K.E., Gardner, W.S., Joyner, A.R., Scott, J.T., 2016b. Mitigating cyanobacterial harmful algal blooms in aquatic ecosystems impacted by climate change and anthropogenic nutrients. Harmful Algae https://doi.org/10.1016/j.hal.2015.09.009.
- Pereyra, J.P.A., D'Agostino, P.M., Mazmouz, R., Woodhouse, J.N., Pickford, R., Jameson, I., Neilan, B.A., 2017. Molecular and morphological survey of saxitoxin-producing cyanobacterium Dolichospermum circinale (Anabaena circinalis) isolated from geographically distinct regions of Australia. Toxicon 138, 68–77. https://doi.org/ 10.1016/j.toxicon.2017.08.006.
- Pestana, C.J., Capelo-Neto, J., Lawton, L., Oliveira, S., Carloto, I., Linhares, H.P., 2019. The effect of water treatment unit processes on cyanobacterial trichome integrity. Sci. Total Environ. 659, 1403–1414. https://doi.org/10.1016/j.scitotenv.2018.12.337.
- Qi, J., Lan, H., Liu, R., Miao, S., Liu, H., Qu, J., 2016. Prechlorination of algae-laden water: the effects of transportation time on cell integrity, algal organic matter release, and chlorinated disinfection byproduct formation. Water Res. 102, 221–228. https://doi.org/ 10.1016/j.watres.2016.06.039.
- Rodríguez, J.J., Mirón, A.S., Camacho, F.G., García, M.C., Belarbi, E.H., Chisti, Y., Grima, E.M., 2009. Causes of shear sensitivity of the toxic dinoflagellate Protoceratium reticulatum. Biotechnol. Prog. 25, 792–800. https://doi.org/10.1002/btpr.161.
- Sheng, H., Liu, H., Wang, C., Guo, H., Liu, Y., Yang, Y., 2012. Analysis of cyanobacteria bloom in the Waihai part of Dianchi Lake, China. Ecol. Inform. https://doi.org/10.1016/j. ecoinf.2012.03.007.
- Sinha, R., Pearson, L.A., Davis, T.W., Burford, M.A., Orr, P.T., Neilan, B.A., 2012. Increased incidence of Cylindrospermopsis raciborskii in temperate zones - is climate change responsible? Water Res. 46, 1408–1419. https://doi.org/10.1016/j.watres.2011.12.019.
- Sun, F., Pei, H.Y., Hu, W.R., Ma, C.X., 2012. The lysis of Microcystis aeruginosa in AlCl 3 coagulation and sedimentation processes. Chem. Eng. J. 193–194, 196–202. https://doi. org/10.1016/j.cej.2012.04.043.
- Szlag, D.C., Sinclair, J.L., Southwell, B., Westrick, J.A., 2015. Cyanobacteria and cyanotoxins occurrence and removal from five high-risk conventional treatment drinking water plants. Toxins 7, 2198–2220. https://doi.org/10.3390/toxins7062198.
- Tornevi, A., Simonsson, M., Forsberg, B., Säve-Söderbergh, M., Toljander, J., 2016. Efficacy of water treatment processes and endemic gastrointestinal illness - a multi-city study in Sweden. Water Res. 102, 263–270. https://doi.org/10.1016/j.watres.2016.06.018.
- Trenberth, K.E., Dai, A., Van Der Schrier, G., Jones, P.D., Barichivich, J., Briffa, K.R., Sheffield, J., 2014. Global warming and changes in drought. Nat. Clim. Chang. 4, 17–22. https:// doi.org/10.1038/nclimate2067.
- Wang, C., Lan, C.Q., 2018. Effects of shear stress on microalgae a review. Biotechnol. Adv. 36, 986–1002.
- Woitzik, D., Weckesser, J., Jurgens, U.J., 1988. Isolation and characterization of cell wall components of the unicellular cyanobacterium synechococcus sp. PCC 6307. Microbiology 134, 619–627. https://doi.org/10.1099/00221287-134-3-619.
- Xie, P., Chen, Y., Ma, J., Zhang, X., Zou, J., Wang, Z., 2016. A mini-review of preoxidation to improve coagulation. Chemosphere 155, 550–563. https://doi.org/10.1016/j. chemosphere.2016.04.003.
- Zamyadi, A., Henderson, R., Stuetz, R., Hofmann, R., Ho, L., Newcombe, G., 2015. Fate of geosmin and 2-methylisoborneol in full-scale water treatment plants. Water Res. 83, 171–183. https://doi.org/10.1016/j.watres.2015.06.038.
- Zarantonello, V., Silva, T.P., Noyma, N.P., Gamalier, J.P., Mello, M.M., Marinho, M.M., Melo, R.C.N., 2018. The cyanobacterium Cylindrospermopsis raciborskii (CYRF-01) responds to environmental stresses with increased vesiculation detected at single-cell resolution. Front. Microbiol. 9, 272. https://doi.org/10.3389/fmicb.2018.00272.
- Zhou, S., Shao, Y., Gao, N., Zhu, S., Li, L., Deng, J., Zhu, M., 2014. Removal of Microcystis aeruginosa by potassium ferrate (VI): impacts on cells integrity, intracellular organic matter release, and disinfection by-products formation. Chem. Eng. J. 251, 304–309. https://doi.org/10.1016/j.cej.2014.04.081.